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Potential decoupling of CO₂ and Hg uptake process by global vegetation in the 21st century

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Mercury (Hg), a potent neurotoxin posing risks to human health, is cycled through vegetation uptake, which is susceptible to climate change impacts. However, the extent and pattern of these impacts are largely unknown, obstructing predictions of Hg's fate in terrestrial ecosystems. Here, we evaluate the effects of climate change on vegetation elemental Hg [Hg(0)] uptake using a state-of-the-art global terrestrial Hg model (CLM5-Hg) that incorporates plant physiology. In a business-as-usual scenario, the terrestrial Hg(0) sink is predicted to decrease by 1870 Mg yr⁻¹ in 2100, that is ~60% lower than the present-day condition. We find a potential decoupling between the trends of CO_2 assimilation and Hg(0) uptake process by vegetation in the 21st century, caused by the decreased stomatal conductance with increasing CO_2 . This implies a substantial influx of Hg into aquatic ecosystems, posing an elevated threat that warrants consideration during the evaluation of the effectiveness of the Minamata Convention.

Mercury (Hg) is a pervasive toxic pollutant causing adverse effects on human health at a global scale¹. Additionally, it endangers ecosystems by bioaccumulating in food chains, affecting biodiversity and disrupting ecological balances². Anthropogenic activities, such as fossil fuel combustion and metal mining, have significantly increased Hg emissions and caused widespread environmental Hg contamination since the industrial era^{3,4}. Vegetation within terrestrial ecosystems can absorb large amounts of atmospheric gaseous elementary Hg [Hg(0)] (2200–3600 Mg year⁻¹), acting as a major sink for the atmosphere in the present-day Hg cycles^{5,6}. Long-living vegetation not only stores present-day Hg but also Hg emitted into the atmosphere decades ago⁷. Climate-related factors, such as rising temperatures and elevated carbon dioxide (eCO₂), are profoundly affecting the growth and physiological processes of vegetation^{8,9}. However, their cascading effects on the terrestrial Hg cycle, especially the uptake of Hg(0) by vegetation, remain unclear. Here, we evaluate this effect by using a coupled climate–land–mercury model running for the twenty-first century.

Previous research has revealed complex impact pathways of climate change on vegetation Hg(0) uptake. The higher temperature was found to boost vegetation's ability to absorb atmospheric Hg(0) in glacier retreat areas¹⁰. Yet, the rising temperatures often lead to localized droughts, which are likely to weaken the Hg sink in terrestrial

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Fig. 1 | **The biogeochemical effect of elevated carbon dioxide on global vegetation uptake of Hg(0).** a Hg(0) vegetation uptake flux at present day, the value represents the gross uptake of atmospheric Hg(0) through both stomatal and cuticular (non-stomatal) processes and does not include the immediate re-emission from foliage. Observations (represented by rhombuses) are obtained from the global vegetation measurements database (see Materials and Methods). **b** Change in Hg(0) vegetation uptake flux caused by the biogeochemical effects of eCO₂ between 2100 and the present day under a business-as-usual scenario (SSP5-8.5). The numbers in the figure represent the global total values.

ecosystems¹¹. Furthermore, the alteration in global precipitation patterns can affect the Hg sink of forests¹². The future increases in vegetation density driven by CO₂ fertilization are expected to enhance the Hg(0) dry deposition velocity^{13,14}. Stomatal uptake of Hg(0) by foliage was found to be inhibited under high vapor pressure deficit (VPD) conditions and proved to be sensitive to extreme climate events¹⁵⁻¹⁷. However, previous studies have not systematically considered the interaction of climate change, vegetation dynamics, and Hg processes, and some have primarily focused only on specific regions^{14,18}.

Therefore, our research aims to exam how vegetation-regulated atmospheric Hg(0) deposition will change under the impact of future climate change. We hypothesize that the climate will influence global vegetation Hg(0) uptake by altering the plant physiology such as the stomatal activities, with CO₂ and other meteorological factors as important driver factors. This study uses the Community Land Model–Hg (CLM5-Hg) within the Community Earth System Model (CESM), which incorporates a dynamic plant growth framework that includes the stomatal uptake process of Hg(0) and various plant functional types (PFTs)¹⁹. This model also encompasses the comprehensive biogeochemical cycling of Hg in terrestrial ecosystems. It is forced by different future climate scenarios throughout the twenty-first century: (i) the Shared Socioeconomic Pathway (SSP) 1-2.6, a.k.a.

the "2°C scenario" representing a sustainability framework; (ii) the SSP3-7.0, representing a medium-high reference within the socioeconomic context of "regional rivalry"; (iii) SSP5-8.5, a.k.a. the "business-as-usual", considered as the worst-case scenario within a high fossil fuel-intensive world²⁰. These future scenarios are compared against the present-day simulation (baseline case). We also include a pre-industrial scenario (ca. 1850) for comparison. Furthermore, we design sensitivity experiments by alternatively changing specific climate-related factors for the SSP5-8.5 scenario while maintaining others consistent with the baseline scenario (Supplementary Table 1). These experiments can diagnose and compare the influence of individual factors on the uptake of Hg(0) by vegetation in terrestrial ecosystems. We consider factors including atmospheric CO₂ concentration (with a focus on biogeochemical effects only), precipitation, temperature, humidity, pressure, radiation, and wind. We keep the anthropogenic Hg emissions and the atmospheric Hg concentrations constant for all scenarios to highlight the impact of climate factors, and remove the effects of land use and land cover change (LULCC) and aerosols (see "Methods").

Results and discussion

Reduced Hg(0) uptake

Our results indicated that, under future climate change scenarios, the biogeochemical effects of elevated CO₂ emerge as the dominant factor influencing vegetation Hg(0) uptake. This uptake represents the gross uptake of atmospheric Hg(0) through both stomatal and cuticular (non-stomatal) processes and does not include the immediate reemission from foliage. In the SSP5-8.5 (business-as-usual) scenario where only CO₂ concentration is altered and other climatological factors kept as present day, the global Hg(0) uptake decreases by 1870 Mg year⁻¹ or 59.6% in 2100, in comparison to the present-day condition of 3138 Mg year⁻¹ (Fig. 1). The most significant changes were simulated in East Asian and the Amazon forests, attributed to their high Hg(0) assimilation compared to other regions¹⁹. The global vegetation Hg(0) uptake was predicted to further decrease by only 88 Mg year⁻¹ while accounting for the changes of all factors in the SSP5-8.5 scenarios (Supplementary Fig. 2). Other climate change factors, such as changing temperature, precipitation, radiation, pressure, and humidity account for a much smaller effect than the biogeochemical effects of eCO₂ alone (Supplementary Figs. 3 and 4b). We also found no significant interaction between the eCO₂ effect and other factors (Supplementary Fig. 5).

We predicted a higher global vegetation Hg(0) uptake for the other two future scenarios with lower CO₂ levels: 2685 Mg year⁻¹ for the SSP1-2.6 (2 °C) scenario (CO₂ = 445.6 p.p.m.) and 1570 Mg year⁻¹ for SSP3-7.0 (regional rivalry) scenario ($CO_2 = 867.2 \text{ p.p.m.}$) vs. 1268 Mg year⁻¹ for SSP5-8.5 (business-as-usual) scenario ($CO_2 = 1135.2 \text{ p.p.m.}$) (Fig. 2). A slight increase in Hg(0) uptake was simulated during the preindustrial era: 3324 Mg year⁻¹ when the global average CO₂ is lower at 288 p.p.m. Unlike the future scenarios, we noted the largest impact is contributed by the lower atmospheric humidity in the pre-industrial era (Supplementary Figs. 3d and 4a). Changes in precipitation and temperature, as well as its interaction with the biogeochemical effects of eCO_2 , significantly affect the uptake of Hg(0) by global vegetation (see Supplementary Figs. 4a and 5a). This suggested that the biogeochemical impact of eCO₂ on Hg(0) uptake has not yet become dominant when compared with other climate change factors in the preindustrial era. When all scenarios were considered together, we observe a continuous decrease in the potential of vegetation to uptake Hg(0) in the future as CO_2 levels increase (Supplementary Fig. 12).

Direct field evidence that integrated the various processes of terrestrial Hg cycling, and was responsive to multiple climate change factors across different spatiotemporal scales, remained scarce. Previous manipulative experiments focusing on CO_2 enrichment have primarily concentrated on the biogeochemical impact of eCO_2 on



Fig. 2 | Global vegetation Hg(0) uptake (bars) via elevated CO_2 under different atmospheric CO_2 concentrations (lines) from the historical emission scenario in 1850 to the Shared Socioeconomic Pathway (SSP) emission scenarios in 2100. SSP1-2.6 represents the lowest scenario, termed the "2 °C scenario," which aims for a sustainable future. SSP3-7.0 represents a moderate scenario, described as a medium-high reference scenario within the socio-economic context of "regional rivalry." SSP5-8.5 represents the highest scenario, also known as "business-as-usual," considered the worst-case scenario in a high fossil fuel-intensive world. Solid lines represent atmospheric CO_2 levels during 1850–2000, shaded lines represent the atmospheric CO_2 levels during 2000–2100 under different scenarios.

vegetation Hg concentrations (Fig. 3a). The CO₂ concentrations were usually enhanced to 360-610 p.p.m. in these experiments, similar to our future scenarios. Data on $\mathrm{Hg}_{\mathrm{veg}}$ (Hg flux or Hg concentration of plant) from six eCO₂ experimental studies were integrated into four eCO₂ conditions based on the levels of increased CO₂ concentration (Supplementary Fig. 10). Under experimental conditions of a 150 and 200 p.p.m. increase in CO2 concentration, Hgveg showed a significant decreasing trend (P < 0.05, Supplementary Fig. 10a). Despite the lack of significant differences at 253 and 360 p.p.m., even showing an opposite trend at 253 p.p.m. (Supplementary Fig. 10b), the overall effect of our meta-analysis (P < 0.01) suggested that eCO₂ had a suppressive effect on vegetation Hg levels (Fig. 3b). The meta-analysis revealed a significant decrease in vegetation Hg levels as a result of eCO₂, showing an average decrease in foliage Hg levels or Hg uptake of 5.87% per 100 p.p.m. increase (95% CI, -6.5% to -5.3%). If translated to the increased CO₂ concentration level under our model's SSP5-8.5 scenario, this change rate reached nearly 50%. The change in terrestrial Hg(0) sink (~60%) simulated by our model was fairly close to the experimentally observed values. Overall, our findings are consistent with existing evidence synthesized from experimental data worldwide, as atmospheric uptake is the major source of Hg in foliage^{6,21}. Indeed, a multitude of studies, including those employing isotopic techniques, have demonstrated that a predominant portion of atmospheric Hg(0)is assimilated into terrestrial ecosystems via vegetation, primarily through stomatal uptake, accounting for over 80% of the total uptake^{22,23}.

We found that this effect varies across different PFTs. The most substantial suppression by eCO_2 was observed in crops (-12.2 ± 1.1% per 100 p.p.m. CO_2), while the suppression in grasses (-5.8 ± 1.1% per



Fig. 3 | The biogeochemical effect of eCO₂ on global vegetation Hg(0) uptake. a Global distribution of eCO₂ experiments included in this meta-analysis. Circles of orange, blue, and red indicate experiments on crops, grass, and tree, respectively. The numbers outside the parentheses represent the increase in CO₂ concentration, while the numbers inside the parentheses indicate the corresponding change rate in Hg_{veg} (based on the four integrated Δ CO₂ levels). **b** The sensitivity of vegetation

Hg in response to eCO_2 in different types of plants across different experimental studies. The triangle represents one unit of eCO_2 (100 p.p.m. increase). Each data point represents the weighted mean values; error bars indicate the 95% confidence intervals. The numbers indicate sample sizes with the number of stars representing significant levels. **c** Modeled annual global Hg(0) uptake by different types of vegetation under present day and future CO_2 levels under the SSP5-8.5 scenario.



Fig. 4 | The biogeochemical effect of eCO_2 on stomatal conductance and leaf area index (LAI) under the SSP5-8.5 scenario. a Changes in global sunlit stomatal conductance (ΔG_{s_s} sun) caused by eCO_2 between present day and 2100. b Changes in global shaded stomatal conductance (ΔG_s sha) caused by the eCO_2 between

100 p.p.m. CO_2) and trees (-5.2 ± 0.3% per 100 p.p.m. CO_2) was relatively similar (Fig. 3b). In our model, similar pattern was simulated. We also extracted the Plant Functional Types (PFTs) corresponding to the species used in the meta-analysis experimental data. We found that the modeled global Hg(0) uptake of trees, crops, and grasses in 2100 all showed a substantial decrease under the SSP5-8.5 scenario (Fig. 3c). Although trees still dominated the reduction of total Hg(0) uptake due to their largest global coverage (Supplementary Fig. 11), crops were the most affected by eCO₂, showing a decline of nearly 60%. Indeed, massive grass species belong to C4 photosynthetic processes of carbon fixation in plants (C4 plants), while trees and crops are predominantly C₃ photosynthetic processes of carbon fixation in plants $(C_3 \text{ plants})^{24}$. The photosynthesis of the former is less limited by ambient atmospheric CO₂ concentrations and subsequently responds less to eCO₂ than the latter²⁵. As C₃ plants dominate global vegetation and account for most of the Earth's current plant life²⁶, an overall significant weakening of global vegetation Hg(0) uptake by the biogeochemical effects of eCO2 was found (Supplementary Fig. 12).

Decoupled CO₂ and Hg

We found the reduced Hg uptake predicted by the model for the future was caused by a decrease in stomatal conductance due to eCO₂. Vegetation uptakes Hg(0) via diffusion through stomatal pores, which is subsequently fixed by foliage^{19,23,27}. Stomatal conductance depends on its aperture and is associated with plant physiological activities²⁸⁻³⁰ (see Eq. 1 in "Methods"). The Medlyn model in our CLM5-Hg model can effectively simulate the response of stomatal conductance to eCO2 (see Supplementary Figs. 7 and 8, details in the model validation section of the Supplementary Information). A 42.6% decrease in the stomatal conductance of global vegetation was projected for the year 2100 with eCO2 under the SSP5-8.5 (business-as-usual) scenario, compared to present-day levels (Fig. 4). Our analysis identified that the sunlit stomatal conductance emerges as the key driving factor influencing the observed reduction in Hg(0) uptake. Specifically, the changes of conductance on the sunlit side of leaves were more consistent with the distribution of vegetation Hg(0) uptake than the shaded sides, indicating the changes in the sunlit side as a primary factor contributing to the diminished global vegetation Hg(0) uptake

between present day and 2100. d Changes in evapotranspiration from canopy

 (ΔET_c) caused by the eCO₂ between present day and 2100.

factor contributing to the diminished global vegetation Hg(0) uptake (Fig. 1). In general, the sunlit side of leaves receives more direct sunlight and heat, prone to stomatal closure³¹. Conversely, the shaded side has a higher stomatal conductance density and is reserved for gas exchange with relatively stable aperture³².

Our study illustrated a complex and extensive feedback mechanism between the terrestrial Hg, water and carbon cycles. The enhancement of photosynthesis is caused by the biogeochemical effect of eCO₂ is accompanied by the loss of plant water content³³. During this process, plants adjust the stomatal aperture to reduce water transpiration and maximize water use efficiency^{34,35}. The Medlyn model in our CLM5-Hg model is consistent with this optimal stomatal theory. With eCO_2 , the CO_2 partial pressure at the leaf surface (C_S) also increased accordingly, leading to enhanced leaf photosynthesis (A_n) (Eq. (1)). However, A_n is constrained by water potential while C_s continues to increase in CLM5²⁹. This resulted in reduced stomatal conductance (g_s) with eCO₂ (Supplementary Fig. 13), leading to decreased evapotranspiration (Fig. 4d). This process induced increased soil water storage, enhancing water use efficiency (Supplementary Fig. 14). This adaptive mechanism ultimately led to a nonlinear relationship between atmospheric CO₂ concentration and stomatal conductance. Stomatal conductance is directly related to vegetation uptake of Hg(0)in our model (Eqs. (1)-(4), for details refer to "Methods"). This relationship explained the gradual reduction in vegetation uptake of Hg(0). We observed this reduction from the pre-industrial and presentday periods to the SSP1-2.6 scenario (Supplementary Fig. 12). Yet, there was a significant decline in the Hg(0) uptake by vegetation from SSP1-2.6 to SSP3-7.0 due to the dramatic increase in atmospheric CO_2 concentration (Fig. 2). Intriguingly, the sensitivity of stomata to eCO₂ diminished gradually under the influence of long-term eCO₂ conditions. This occurred because guard cells and mesophyll tissues, which mediate stomatal movements, lead to decreases in stomatal aperture and size, culminating in physiological adaptation to higher concentrations^{36,37}. Consequently, this resulted in a less pronounced decline from the SSP3-7.0 to SSP5-8.5 scenarios (Fig. 2).

A tight coupling between carbon and Hg in terrestrial ecosystems has been observed as a paradigm over the past two decades^{10,38}. Conventionally, it has been postulated that rising atmospheric CO_2 levels would increase vegetation's photosynthesis rate, leading to the beneficial impact on plant growth, known as the CO₂ fertilization effect¹¹. This effect is believed to enhance the concurrent absorption of both CO₂ and Hg(0), as suggested by Jiskra et al.³⁹, Obrist⁴⁰, and Schaefer et al.⁴¹. For example, atmospheric Hg and CO₂ have similar seasonal fluctuation patterns in both hemispheres, regulated by vegetation photosynthetic activity. An increase in terrestrial net primary production has been also speculated to contribute to a diminishing trend in atmospheric Hg(0) levels in the Northern Hemisphere over the past two decades^{39,42,43}. Contrarily, we predicted a potential decoupling between the trends of CO_2 assimilation and Hg(0) uptake process by vegetation in the twenty-first century, when considering the dynamic response of vegetation physiological activities to climate change. The CLM5 model projected an increased greening of vegetation in many regions in the twenty-first century resulting from eCO₂ (a.k.a. fertilization effect), evidenced by the increased leaf area index (LAI) in the northern mid-to-low latitudes and certain regions of the Southern Hemisphere (Fig. 4c). The increase in photosynthesis can simultaneously induce a state of water deficit and nutrient saturation within the plant's internal environment⁴⁴. Therefore, under climate change, the increase in vegetation LAI may only represent an increase in leaf density or even stomatal numbers, but stomatal conductance may not necessarily increase accordingly. However, our model suggested a discernible decrease in the flux of Hg(0) uptake by vegetation in these areas (Fig. 1), reflecting the differences in CO₂ and Hg element during plant physiological processes especially those related to water dynamics in terrestrial ecosystems45.

Uncertainties

We noted significant uncertainties in our model results. First, various climate forcing may introduce future uncertainties in CLM5 simulations, such as underestimating the phenology and photosynthesis of future plants⁴⁶. This occurred in part because the anomaly forcing method assumes that future changes (anomalies) can overlay present-day variability. Different sub-monthly variations may not accurately represent all facets of future climate changes, particularly in the presence of non-linear interactions or crossed thresholds^{29,47}. Additionally, it is important to note that some data sources in our meta-analysis originate from seedling experiments. There were inherent physiological and morphological differences between young seedlings and fully mature plants, which could potentially influence the study's outcomes. However, young seedlings often exhibit more pronounced responses to environmental changes, making them suitable for detecting initial patterns and mechanisms in plant response to elevated CO2. Additionally, there will be differences associated with different plant species. Thus, we suggest that future research should focus on this aspect, aiming to bridge the knowledge gap by including experiments across various growth stages and more species.

Given that vegetative stomatal absorption is a key mechanism in our model, we conducted an uncertainty analysis for the parameterization of g_s . We included the sensitivity analysis of five ecologically significant parameters (Medlyn_slope, slatop, leafCN, psi50, and stem_leaf) under four levels of perturbation (Supplementary Table 3). We found that the global vegetation Hg(0) uptake and g_s range 1160–1370 Mg year⁻¹ and 29,600–41,900 µmol H₂O m⁻² s⁻¹, respectively, with an uncertainty ratio of 17% and 21%, respectively (Supplementary Fig. 15). The coefficient of variation (CV, defined as the relative degree of change in the model output compared to the proportion of parameter changes) can reflect the magnitude of an individual parameter's contribution to uncertainty⁴⁸. The sensitivity analysis revealed that the parameter "Medlyn slope" has the highest CV (1.01 and 1.32 for vegetation Hg(0) uptake and g_s , respectively) (Supplementary Fig. 16 and Supplementary Table 4). In the Medlyn model within CLM5, the Medlyn slope, denoted as " g_1 ," plays a crucial role in controlling how stomata respond to CO₂ levels. It does this by determining the extent to which stomata open, based on the assimilation capacity, CO₂ concentration, and VPD⁴⁷. However, the CLM5 model does not differentiate this parameter for different climate types, which induces relatively large uncertainties $(6 \pm 1.2\%)$. Additionally, the stomatal conductance simulated by our model is slightly lower than the observed values (Supplementary Fig. 9), which could be caused by the uncertainty associated with this parameter. Indeed, Kauwe et al.⁴⁹ found a ~30% reduction of the annual transpiration fluxes after better constraining this parameter. This implies that the actual future decrease in Hg(0) uptake could potentially be even higher. More vegetation physiological parameters and Hg observations for different PFTs are thus needed to better constrain our model. There are also likely interaction effects among parameters.

There are still considerable uncertainties regarding the model representation of the land-atmosphere exchange of Hg at present day, which serves as a baseline for our prediction for the future. The atmospheric Hg concentrations and deposition were specified as a boundary condition, not yet dynamically modeled in a two-way coupled fashion. The feedback between land Hg emissions and their atmospheric abundance and subsequent deposition onto the land are also not considered. Although our current framework can well diagnose the direct impact of changing climate on these exchange fluxes, an online land-atmosphere coupled model will be needed to reveal a more comprehensive and accurate changes in global Hg budget in future works. Additionally, current isotopic evidence indicates that the photoreduction process is related to the re-emission of Hg(0) by vegetation leaves, with this re-emission ratio reaching nearly 30% in subtropical forest areas⁵⁰. However, for the majority of other regions worldwide, we lack sufficient observational data to make estimates. In our model, we have only used median values as the reduction parameter¹⁹. Therefore, in future research, we need to utilize more measured data to refine our parameterization scheme. Our model also did not consider the absorption of Hg from underground root systems and root secretions⁵¹. Indeed, Hg is hard to enter the plants via the root. as most previous studies have shown⁵²⁻⁵⁴. Meanwhile, our model did not account for the translocation of Hg among plant tissue organs. A recent study suggested that a significant proportion of Hg in roots may originate from absorption by leaves and subsequent translocation, with an estimation of up to 300 Mg year⁻¹ of atmospheric Hg° stored in roots⁵⁵, but the specific migration and distribution mechanisms are still unclear. Furthermore, the model simplified the soil Hg processes, following GTMM (Global Terrestrial Mercury Model)³⁸, by focusing mainly on the microbial reduction process. It did not account for other processes like the radiative transfer in soil, photo-reduction, and other abiotic reduction processes^{56,57}. These processes also have a potential influence on the amount of Hg(0) uptake by the vegetation, and could be incorporated in our model when more data is available.

In our CLM5-Hg model, throughfall primarily originates from the washing off of atmospheric divalent mercury (sum of the Hg(II) dry deposition onto the canopy surface and the Hg(II) wet deposition that has not been reduced)^{38,58}. Recent studies indicated that epiphytic vegetation on canopies absorbs atmospheric Hg(0) and decomposes into humus, adhering to tree trunks and canopies, where mercury is subsequently washed into throughfall by precipitation⁵⁹. Additionally, research indicated that the temporal scale and frequency of sampling for throughfall mercury measurements can impact the accuracy of their estimates⁶⁰. Therefore, our model has limitations in this part, and more extensive experimental data covering broader spatiotemporal scales is needed to further constrain the model (e.g., flux measurements or isotope compositions). The anthropogenic and legacy Hg emissions from land and ocean also remained unchanged in this study.

Additional uncertainties also aroused from our current understanding of the biogeochemical response of plants to eCO₂ and their possible adaptability to long-term changes²⁴. Overall, these uncertainties necessitate further calibration of the model when more data is available and can be effectively addressed as scientific knowledge evolves. The model should be interpreted as a diagnostic tool designed to unveil the influence of individual factors. It serves as a foundation for a more realistic and comprehensive prediction that takes into account factors such as the connection between future Hg and greenhouse gas emissions^{1,61}, the enhanced soil microbial activity⁴¹, ocean warming and acidification⁶², and amid many others^{12,63}.

Implications

We found that, in the climate change scenario, the atmospheric Hg(0) uptake by terrestrial vegetation in 2100 will be likely to decrease by more than half compared to present-day conditions. The atmospheric CO₂ concentration is an important factor that will impact vegetation Hg uptake in the future. The continuous increase in CO₂ concentration will lead to a warming effect, which alters global precipitation patterns^{64,65}. This could potentially increase VPD and cause drought in many regions, consequently affecting the Hg processes in terrestrial ecosystems. In addition, the composition of global plant communities could be modified over extended time scales⁶⁶. For instance, persistent severe drought could lead to widespread vegetation mortality and shifts in the composition of tropical forest tree⁶⁷. Consequently, this affects the distribution pattern and magnitude of vegetation Hg flux⁶⁸.

Our findings revealed a suppression of atmospheric Hg(0) uptake by plants across most regions in the twenty-first century due to reduced stomatal conductance in vegetation caused by increased CO₂. With climate change, the bypassing of atmospheric Hg(0) sequestration by plants and the deposition of foliar Hg to the soil lead to increasing concentrations in the atmosphere. This Hg can then be converted to HgII, which is deposited in aquatic ecosystems and can subsequently be methylated^{6,18,54}. Furthermore, these inorganic Hg compounds are transformed into methylmercury by microbes. This process leads to the enrichment of methylmercury in riverine and marine food chains. As a result, a substantial threat to human health arises through the consumption of inland aquatic animals and seafood, including commercial fish⁶⁹⁻⁷¹. These processes coincide with the changes in land use/land cover, such as the potential shift from Amazon rainforest to savannah, which also decreases the land Hg sink and contributes to an additional movement of Hg into the ocean¹⁸, construing an additional climate change penalty via Hg cycles. Furthermore, although the impact of anthropogenic source emissions was not the focus of this study, some estimations indicated that global anthropogenic emissions of Hg will increase in the forthcoming decades under the current legislative scenario^{61,72,73}. Therefore, under future climate change scenarios, it is possible that there will be a greater threat to human health. We did not consider the impacts LULCC in this study. Under global warming, vegetative succession following melting and increased precipitation intensity is likely to lead to an increase in vegetative biomass and, consequently, an increase in Hg(0) uptake by vegetation¹⁰. Indeed, the interactive effects of climate change combined with changes in LULCC worth further examination.

The terrestrial ecosystem, recognized as a significant Hg sink, may face disruptions under future climate change scenarios, particularly with rising atmospheric CO₂ concentrations^{5,74}. Therefore, it becomes crucial to comprehensively consider the tight coupling among Hg, CO₂, and water cycles when assessing the effectiveness of the Minamata Convention within the context of climate change. From the perspective of the global Hg cycle, considering only the air-land exchange process is insufficient to achieve global mass balance. Both anthropogenic releases and the air-ocean exchange of Hg can potentially affect atmospheric Hg levels, thus influencing the air-land exchange process. Therefore, future research should aim to further incorporate time-varying anthropogenic emissions and develop a fully coupled land-atmosphere-ocean global Hg model within CESM2. This would enable a comprehensive understanding of the complete pathway of Hg from emission to deposition.

Methods

CLM5-Hg

We applied a state-of-the-art global terrestrial Hg model to explore the impact of climate change on global vegetation Hg(0) uptake. We used the CLM5-Hg model, which comprehensively contains the biogeophysical and biogeochemical processes that control terrestrial Hg cycling (Supplementary Fig. 1). CLM5-Hg was tested against observational field data by Yuan et al.¹⁹. The model simulates the migration, transformation, accumulation, and emission processes of Hg in terrestrial ecosystems. This includes processes such as the stomatal and non-stomatal uptake of Hg(0) in leaves, the throughfall of divalent Hg [Hg(II)], and the formation of litter and soil Hg. It also includes processes such as the leaching of soil Hg, photo-reduction, as well as microbial decomposition, thermal evaporation, and emissions from wildfires.

The CLM5 model represents its surface heterogeneity with multilayer nested grid cells. The first layer of the sub-grid consists of land units, including five types: vegetation, lakes, cities, glaciers, and crops. The second layer of the sub-grid represents soil columns, indicating the state changes of the soil within the same land unit. The third layer of the sub-grid is plant functional types (PFTs) with different biogeochemical processes. Sub-grids within the same model grid use the same atmospheric forcing dataset, but the diagnostic variables for each sub-grid are simulated independently. The vegetation covering the land surface is composed of 16 different PFTs (temperate-needleleaf evergreen tree, boreal-needleleaf evergreen tree, borealneedleleaf deciduous tree, tropical-broadleaf evergreen tree, temperate-broadleaf evergreen tree, tropical-broadleaf deciduous tree, temperate-broadleaf deciduous tree, boreal-broadleaf deciduous tree, temperate- broadleaf evergreen shrub, temperate-Broadleaf deciduous shrub, boreal-broadleaf deciduous shrub, C₃ arctic grass. C₃ grass, C₄ grass, and crops)¹⁹. Their differences in leaves and stems determine the uniqueness of different vegetation in reflection, transmittance, and solar radiation absorption. Root distribution parameters control soil moisture absorption, aerodynamic parameters determine thermal resistance, moisture, and momentum transfer, and photosynthetic parameters determine stomatal resistance, photosynthesis, and evapotranspiration processes. All PFTs are divided into three different types of phenology: perennial evergreen types, seasonal deciduous types determined by temperature and daylight length, and multi-seasonal stress deciduous types determined by temperature and soil moisture²⁹.

We used the offline version of CLM5 (0.90° latitude ×1.25° longitude) with coupled biogeochemical cycles (BGC), which is forced by the dataset of Global Soil Wetness Project (GSWPS), a 3-hourly 0.5° global forcing product based on 20th Century Reanalysis version. The biogeochemical model was run for 200 years, and the CLM5-Hg was run for 10 years in each scenario simulation (1841–1850 for the preindustrial era, 1991–2000 for the present day, and 2091-2100 for the future). The results for the last year of each simulation were used for data analysis, as this time point reached a steady state, as indicated by two representative variables (see Supplementary Fig. 17). We used the simulated atmospheric Hg(0) concentrations and the dry and wet deposition fluxes of Hg(II) from the CAM6-Chem model as its upper boundary conditions⁷⁵. To isolate the impact of climate change, anthropogenic Hg emissions were maintained at the present-day level⁷⁶.

Stomatal model

In CLM5-Hg, maximum stomatal conductance was obtained from the Medlyn "empirical-optimal" conductance $model^{30}$. This stomatal model calculates stomatal conductance (g_s) based on net leaf

photosynthesis, the CO_2 concentration at the leaf surface, and the VPD. The stomatal resistance of the leaf is:

$$\frac{1}{r_s} = g_s = g_0 + 1.6 \left(1 + \frac{g_1}{\sqrt{D}} \right) \frac{A_n}{C_s / P_{atm}}$$
(1)

where r_s is the stomatal resistance, g_0 is the minimum stomatal conductance, A_n is leaf net photosynthesis, C_s is the CO_2 partial pressure at the leaf surface, P_{atm} is the atmospheric pressure, and *D* is the VPD at the leaf surface. The value of g_1 depends on the PFTs following the CABLE model⁴⁹. The model further have corrected r_s by partitioning to sunlit and shaded side leaf stomatal resistance and the condition of snow cover²⁹:

$$R_{s} = \left(\frac{f_{sun} \times elai}{r_{s}^{sun}} + \frac{(1 - f_{sun}) \times elai}{r_{s}^{sha}}\right)^{-1}$$
(2)

where R_s represents the adjusted stomatal resistance, f_{sun} is the sunlit fraction of canopy, elai represents one-sided LAI buried by snow, r_s^{sun} is the sunlit leaf stomatal resistance, and r_s^{sha} is the shaded leaf stomatal resistance.

The dry deposition flux is used to calculate the absorption of atmospheric Hg(0) by global vegetation⁷⁷:

$$F_d(z) = V_d C_{(z)} \tag{3}$$

where $F_d(Z)$ represents the Hg(0) dry deposition flux at height z, $C_{(z)}$ is the atmospheric Hg(0) concentration, and v_d is the dry deposition velocity calculated following the Wesely scheme⁷⁸:

$$V_d = \frac{1}{R_a + R_b + R_c} \tag{4}$$

where R_a is the aerodynamic resistance between a specific height and the surface, R_b is the quasi-laminar sublayer resistance, and R_c is the bulk surface resistance. R_c is mainly determined by the adjusted stomatal resistance (R_s).

Experiment design

We simulated the terrestrial Hg cycling that is representative of the present-day conditions to serve as the baseline, and 1850 as the preindustrial climate condition. We selected three different SSPs to represent three different future CO2 emission scenarios in 2100: SSP1-2.6 represents the lowest, SSP5-8.5 represents the highest, and SSP3-7.0 represents a moderate scenario. To better unravel the effects of individual factors on global vegetation uptake of Hg(0), a suite of simulations was performed in the pre-industrial era and the hypothesized twenty-first century under the SSP5-8.5 scenario. To mitigate high computational costs, we adopted an alternative approach known as "anomaly forcing" for land-only simulations in CLM5 in line with future climate projections, enabling the generation of climate data to drive CLM5⁷⁹. This method used data from a fully coupled simulation to produce monthly changes in near-surface atmospheric states and fluxes, relative to current conditions⁷⁹. We isolated the effects of individual factors by sequentially altering only one variable to reflect future states while maintaining other variables in line with present-day conditions. To eliminate the impact of land use and land cover changes, as well as aerosols, we preprocessed the vegetation patterns and aerosol deposition in all simulation scenarios using the specified file provided by CLM5. These climate change factors specifically include atmospheric CO₂ concentration (limited to its biogeochemical effects and its climate effects are reflected in other factors), precipitation, radiation, meteorological factors such as temperature, humidity, pressure, and wind. In the "all" scenario, all these factors were modified simultaneously to simulate the overall effect. Additionally, we examined the interactions between the biogeochemical effect of CO_2 and temperature and precipitation. We also have intentionally maintained constant levels of anthropogenic emissions and atmospheric Hg concentrations across all scenarios. This methodological choice was made to isolate and underscore the influence of climatic variables on Hg dynamics. By controlling for anthropogenic inputs, we aim to provide a clear assessment of how climate change alone can affect the biogeochemical cycling of Hg.

Data and meta-analysis

We utilized a global vegetation Hg(0) flux dataset to constrain and validate our simulated global vegetation assimilation of Hg(0) at present day. This database contained 79 publications with measurements ranging from 1987 to 2020¹⁸. These data were from 37 different measurement sites across the world covering the East Asia, Western Europe, and North America regions. The modeled Hg(0) vegetation uptake flux was constrained by the 60 individual data points and agrees well with the observations ($r^2 = 0.38$)¹⁹. The datasets of global stomatal conductance were used to validate the accuracy of stomatal uptake process by CLM5-Hg³⁷.

We used meta-analysis to assess the biogeochemical effects of experimentally elevated CO_2 (eCO₂) on foliar Hg concentration. We searched for journal articles using the ISI Web of Science with the following keyword combinations: (elevated CO_2 concentration OR CO_2 enrichment OR increasing CO_2 concentration) AND (mercury OR mercury concentration OR mercury uptake) AND (tree OR grass OR plants OR vegetation OR leaf OR leaves) from 1990 to 2023. Papers have to meet the following criteria to be included in our dataset: (i) eCO₂ experiments were conducted in terrestrial ecosystems; (ii) initial environmental factors in control plots were the same as those in eCO₂ plots; (iii) at least two CO_2 concentration from six studies^{80–85}. To make sure we include all important studies, we did another search using Google Scholar and sort the studies based on their relevance.

The weighted mean response ratio (InRR) is employed to analyze the treatment effect on vegetation mercury (Hg)^{37,86}, the effect size is estimated as:

$$\ln RR = \ln \left(\frac{\overline{x}_{eCO_2}}{\overline{x}_{aCO_2}} \right)$$
(5)

where \overline{x}_{eCO_2} and \overline{x}_{aCO_2} represent the mean of the elevated CO₂ level and ambient CO₂ level, respectively.

The individual observations are assigned weights calculated from the experimental replications:

weight =
$$\left(\frac{n_{eCO_2} \times n_{aCO_2}}{n_{eCO_2} + n_{aCO_2}}\right)$$
 (6)

where n_{eCO_2} and n_{aCO_2} indicate the numbers of replications at elevated CO₂ level and ambient CO₂ level, respectively.

Finally, InRR is transformed to percentage change (%) as:

$$RR\% = (\exp(\ln RR) - 1) \times 100 \tag{7}$$

The natural log-transformed foliar Hg concentration or flux (Hg_{veg}) sensitivity (InSens) is calculated as:

$$\ln Sens = \frac{\ln RR}{\Delta}$$
(8)

where lnRR represents the natural log-transformed response ration, and the Δ is the magnitude of eCO₂ (per 100 p.p.m. increase). The weighted means of lnSens and percentage sensitivity are calculated using equations similar to those presented in Eqs. (6) and (7) above.

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A negative effect size indicate a decline in the Hg_{veg} (response variable) for the treatment plots compared to the control plots. A variable is considered significantly different between the treatment and control plots (P < 0.05) if the 95% confidence intervals (CI) of the effect size for that variable does not overlap with zero. Differences between subgroups are deemed significant if their CIs do not overlap. Furthermore, due to the limitations of the data, it is not feasible to specifically subdivide PFTs for a direct one-to-one comparison between the model output and the observational data. Therefore, we have roughly categorized these studies into three subgroups: tree, grass, and crop (Supplementary Table 2). This categorization aims to observe the impact of eCO_2 on plants from different PFTs, enabling comparison with our model results.

Uncertainty analysis

We conducted an uncertainty analysis on the Hg(0) uptake and g_s using a perturbation experiment. This was based on the process of stomatal uptake, which is a crucial part in the assimilation of atmospheric Hg(0) by global vegetation⁵⁵. Our focus was on a set of five plant physiology-related parameters within the stomatal conductance model (Medlyn model in CLM5), selected due to their mechanical impacts on responses to elevated CO₂ levels and their significance in representing important ecological processes in vegetation⁴⁷ (Supplementary Table 3). The Medlyn_slope is the slope of the Medlyn model (Eq. 1), where the slope parameter dictates the extent of stomatal opening based on a given mix of assimilation capacity, CO2 concentration, and VPD. Slatop measures the leaf area per gram of leaf biomass. Higher SLA values, indicating thinner and more efficient leaves, lead to a greater LAI from the same biomass. Leaf cn denotes the ideal leaf carbon to nitrogen (C:N) ratio, with higher leaf nitrogen enhancing photosynthesis but increasing respiration costs. Psi50 is the water potential at which there is a 50% loss of conductivity, a hydraulic trait of plants that effectively reflects the vegetation's state in response to water-deficient conditions. Stem leaf determines the biomass distribution between stem and leaves. As this ratio increases, it diminishes the achievable LAI per unit of carbon and nitrogen dedicated to growth, while concurrently increasing the equilibrium woody biomass. These parameters were selected from the broader set of CLM5 parameters through a method primarily guided by the model's structure and thorough iterative testing during CLM5's development⁴⁷. To simplify understanding and pinpointing model behavior changes due to parameter variations, we employed one-at-a-time (OAAT) perturbations from the default settings. This approach, unlike a comprehensive global parameter sensitivity analysis, enables easier visualization and interpretation of results. We assessed the sensitivity of simulations in the CLM5-Hg model under a standard scenario, testing four levels for each parameter as outlined in Supplementary Table 3, resulting in a total of 20 distinct physical perturbation ensembles. In addition, we also considered a condition in which all parameters are altered simultaneously at each level. We estimated the relative uncertainty in Hg(0) uptake and stomatal conductance using the CVs, defined as the relative degree of change in the model output compared to the proportion of parameter changes⁴⁸.

Model validation

We curated a dataset within the CO_2 concentration ranges projected by our model under three SSP scenarios, which included 462 paired observations of treated versus control groups for various species across different vegetation biomes (Supplementary Figs. 6 and 7). First, the comparison results under the current scenario indicated that the simulation can captures the global pattern of stomatal conductance (r > 0.5, Supplementary Fig. 8). Second, we performed pointby-point validations against matched observational data for projected eCO_2 concentrations in each scenario, achieving an r value of 0.50 (Supplementary Fig. 9). Our previous research has shown that the CLM5-Hg model can capture the global distribution of vegetative Hg and litter Hg concentrations well (r > 0.6) via the Hg(0) dataset and a large number of observational datasets related to vegetative Hg tissue concentration¹⁹.

Reporting summary

Further information on research design is available in the Nature Portfolio Reporting Summary linked to this article.

Data availability

All data generated or analyzed are available in the main text, the Supplementary information, and the research group website: https://www.ebmg.online/mercury. GSWP3 climate dataset for CESM2: https://svnccsm-inputdata.cgd.ucar.edu/trunk/inputdata/atm/datm7/. The data from the uncertainty analysis and the sensitivity experiment of model generated in this study are provided in the Supplementary Information/Source data file. Source data are provided with this paper.

Code availability

All core CLM5-Hg model code is available at the research group website: https://www.ebmg.online/mercury. The CESM2 code: https:// github.com/ESCOMP/CESM. The CLM5 code: https://github.com/ ESCOMP/CTSM/tree/master/src.

References

- 1. Zhang, Y. et al. Global health effects of future atmospheric mercury emissions. *Nat. Commun.* **12**, 3035 (2021).
- Eagles-Smith, C. A. et al. Mercury in western North America: a synthesis of environmental contamination, fluxes, bioaccumulation, and risk to fish and wildlife. *Sci. Total Environ.* 568, 1213–1226 (2016).
- Amos, H. M., Jacob, D. J., Streets, D. G. & Sunderland, E. M. Legacy impacts of all-time anthropogenic emissions on the global mercury cycle. *Glob. Biogeochem. Cycles* 27, 410–421 (2013).
- 4. Gerson, J. R. et al. Amazon forests capture high levels of atmospheric mercury pollution from artisanal gold mining. *Nat. Commun.* **13**, 1–10 (2022).
- 5. Daniel, O. et al. Previously unaccounted atmospheric mercury deposition in a midlatitude deciduous forest. *Proc. Natl Acad. Sci. USA* **118**, e2105477118 (2021).
- Zhou, J., Obrist, D., Dastoor, A., Jiskra, M. & Ryjkov, A. Vegetation uptake of mercury and impacts on global cycling. *Nat. Rev. Earth Environ.* 2, 269–284 (2021).
- 7. Bargagli, R. Moss and lichen biomonitoring of atmospheric mercury: a review. *Sci. Total Environ.* **572**, 216–231 (2016).
- Higgins, S. I., Conradi, T. & Muhoko, E. Shifts in vegetation activity of terrestrial ecosystems attributable to climate trends. *Nat. Geosci.* 16, 147–153 (2023).
- Seddon, A. W. R., Macias-Fauria, M., Long, P. R., Benz, D. & Willis, K. J. Sensitivity of global terrestrial ecosystems to climate variability. *Nature* 531, 229–232 (2016).
- Wang, X. et al. Global warming accelerates uptake of atmospheric mercury in regions experiencing glacier retreat. *Proc. Natl Acad. Sci. USA* 117, 2049–2055 (2020).
- Liu, Y. et al. Field-experiment constraints on the enhancement of the terrestrial carbon sink by CO₂ fertilization. *Nat. Geosci.* 12, 809–814 (2019).
- 12. Sonke, J. E. et al. Global change effects on biogeochemical mercury cycling. *Ambio* **52**, 853–876 (2023).
- Wu, S., Mickley, L. J., Kaplan, J. O. & Jacob, D. J. Impacts of changes in land use and land cover on atmospheric chemistry and air quality over the 21st century. *Atmos. Chem. Phys.* 12, 1597–1609 (2012).
- Zhang, H., Holmes, C. D. & Wu, S. Impacts of changes in climate, land use and land cover on atmospheric mercury. *Atmos. Environ.* 141, 230–244 (2016).

- Wohlgemuth, L. et al. Physiological and climate controls on foliar mercury uptake by European tree species. *Biogeosciences* 19, 1335–1353 (2022).
- 16. Wohlgemuth, L., Feinberg, A., Buras, A. & Jiskra, M. A spatial assessment of current and future foliar Hg uptake fluxes across European forests. *Glob. Biogeochem. Cycles* **37**, e2023GB007833 (2023).
- Damour, G., Simonneau, T., Cochard, H. & Urban, L. An overview of models of stomatal conductance at the leaf level. *Plant Cell Environ*. 33, 1419–1438 (2010).
- Feinberg, A., Dlamini, T., Jiskra, M., Shah, V. & Selin, N. E. Evaluating atmospheric mercury (Hg) uptake by vegetation in a chemistrytransport model. *Environ. Sci. Process. Impacts* 24, 1303–1318 (2022).
- Yuan, T. et al. Buffering effect of global vegetation on the air-land exchange of mercury: insights from a novel terrestrial mercury model based on CESM2-CLM5. *Environ. Int.* **174**, 107904 (2023).
- Meinshausen, M. et al. The shared socio-economic pathway (SSP) greenhouse gas concentrations and their extensions to 2500. *Geosci. Model Dev.* 13, 3571–3605 (2020).
- Fu, X. et al. Depletion of atmospheric gaseous elemental mercury by plant uptake at Mt. Changbai, Northeast China. *Atmos. Chem. Phys.* 16, 12861–12873 (2016).
- Laacouri, A., Nater, E. A. & Kolka, R. K. Distribution and uptake dynamics of mercury in leaves of common deciduous tree species in Minnesota, U.S.A. *Environ. Sci. Technol.* 47, 10462–10470 (2013).
- Liu, Y. et al. Understanding foliar accumulation of atmospheric Hg in terrestrial vegetation: progress and challenges. *Crit. Rev. Environ. Sci. Technol.* 52, 4331–4352 (2021).
- Reich, P. B., Hobbie, S. E., Lee, T. D. & Pastore, M. A. Unexpected reversal of C3 versus C4 grass response to elevated CO₂ during a 20-year field experiment. *Science* **360**, 317–320 (2018).
- SAGE, R. F. & KUBIEN, D. S. The temperature response of C3 and C4 photosynthesis. *Plant Cell Environ.* **30**, 1086–1106 (2007).
- Still, C. J., Berry, J. A., Collatz, G. J. & DeFries, R. S. Global distribution of C3 and C4 vegetation: carbon cycle implications. *Glob. Biogeochem. Cycles* 17, 6–14 (2003).
- 27. Obrist, D. et al. Tundra uptake of atmospheric elemental mercury drives Arctic mercury pollution. *Nature* **547**, 201–204 (2017).
- Franks, P. J. et al. Sensitivity of plants to changing atmospheric CO₂ concentration: from the geological past to the next century. *N. Phytol.* **197**, 1077–1094 (2013).
- 29. Lawrence, D. M. et al. The community land model version 5: description of new features, benchmarking, and impact of forcing uncertainty. *J. Adv. Model. Earth Syst.* **11**, 4245–4287 (2019).
- Medlyn, B. E. et al. Reconciling the optimal and empirical approaches to modelling stomatal conductance. *Glob. Change Biol.* 18, 3476 (2012).
- Savvides, A., Fanourakis, D. & van Ieperen, W. Co-ordination of hydraulic and stomatal conductances across light qualities in cucumber leaves. J. Exp. Bot. 63, 1135–1143 (2012).
- 32. Haworth, M. et al. Allocation of the epidermis to stomata relates to stomatal physiological control: stomatal factors involved in the evolutionary diversification of the angiosperms and development of amphistomaty. *Environ. Exp. Bot.* **151**, 55–63 (2018).
- Katul, G. G., Oren, R., Manzoni, S., Higgins, C. & Parlange, M. B. Evapotranspiration: a process driving mass transport and energy exchange in the soil-plant-atmosphere-climate system. *Rev. Geophys.* https://doi.org/10.1029/2011RG000366 (2012).
- Gardner, A. et al. Optimal stomatal theory predicts CO₂ responses of stomatal conductance in both gymnosperm and angiosperm trees. *N. Phytol.* **237**, 1229–1241 (2023).
- Hsiao, J., Swann, A. L. S. & Kim, S.-H. Maize yield under a changing climate: the hidden role of vapor pressure deficit. *Agric. For. Meteorol.* 279, 107692 (2019).

- Engineer, C. B. et al. CO₂ sensing and CO₂ regulation of stomatal conductance: advances and open questions. *Trends Plant Sci.* 21, 16–30 (2016).
- 37. Liang, X. et al. Stomatal responses of terrestrial plants to global change. *Nat. Commun.* **14**, 2188 (2023).
- Smith-Downey, N. V., Sunderland, E. M. & Jacob, D. J. Anthropogenic impacts on global storage and emissions of mercury from terrestrial soils: insights from a new global model. *J. Geophys. Res. Biogeosci.* https://doi.org/10.1029/2009JG001124 (2010).
- 39. Jiskra, M. et al. A vegetation control on seasonal variations in global atmospheric mercury concentrations. *Nat. Geosci.* **11**, 244–250 (2018).
- 40. Obrist, D. Atmospheric mercury pollution due to losses of terrestrial carbon pools? *Biogeochemistry* **85**, 119–123 (2007).
- 41. Schaefer, K. et al. Potential impacts of mercury released from thawing permafrost. *Nat. Commun.* **11**, 4650 (2020).
- Fu, X. et al. Significant seasonal variations in isotopic composition of atmospheric total gaseous mercury at forest sites in China caused by vegetation and mercury sources. *Environ. Sci. Technol.* 53, 13748–13756 (2019).
- St. Louis, V. L. et al. Atmospheric concentrations and wet/dry loadings of mercury at the remote Experimental Lakes Area, Northwestern Ontario, Canada. *Environ. Sci. Technol.* 53, 8017–8026 (2019).
- 44. Victoria, G., Jean-Iouis, D. & François, G. Water deficit and nitrogen nutrition of crops. A review. *Agron. Sustain. Dev.* **30**, 529–544 (2010).
- Wang, X., Yuan, W., Lin, C.-J. & Feng, X. Mercury cycling and isotopic fractionation in global forests. *Crit. Rev. Environ. Sci. Technol.* 52, 3763–3786 (2021).
- Lu, Y. & Yang, X. Using the anomaly forcing Community Land Model (CLM 4.5) for crop yield projections. *Geosci. Model Dev.* 14, 1253–1265 (2021).
- Fisher, R. A. et al. Parametric controls on vegetation responses to biogeochemical forcing in the CLM5. J. Adv. Model. Earth Syst. 11, 2879–2895 (2019).
- Cui, J. et al. Nitrogen cycles in global croplands altered by elevated CO₂. Nat. Sustain. 6, 1166–1176 (2023).
- De Kauwe, M. G. et al. A test of an optimal stomatal conductance scheme within the CABLE land surface model. *Geosci. Model Dev.* 8, 431–452 (2015).
- 50. Yuan, W. et al. Stable isotope evidence shows re-emission of elemental mercury vapor occurring after reductive loss from foliage. *Environ. Sci. Technol.* **53**, 651–660 (2019).
- Keuper, F. et al. Carbon loss from northern circumpolar permafrost soils amplified by rhizosphere priming. *Nat. Geosci.* 13, 560–565 (2020).
- Chiarantini, L. et al. Black pine (*Pinus nigra*) barks as biomonitors of airborne mercury pollution. *Sci. Total Environ.* 569–570, 105–113 (2016).
- Siwik, E. I. H., Campbell, L. M. & Mierle, G. Distribution and trends of mercury in deciduous tree cores. *Environ. Pollut.* 158, 2067–2073 (2010).
- 54. Arnold, J., Gustin, M. S. & Weisberg, P. J. Evidence for nonstomatal uptake of Hg by aspen and translocation of Hg from foliage to tree rings in Austrian pine. *Environ. Sci. Technol.* **52**, 1174–1182 (2018).
- 55. Zhou, J. & Obrist, D. Global mercury assimilation by vegetation. *Environ. Sci. Technol.* **55**, 14245–14257 (2021).
- Fritsche, J. et al. Elemental mercury fluxes over a sub-alpine grassland determined with two micrometeorological methods. *Atmos. Environ.* 42, 2922–2933 (2008).
- Wang, X. et al. Emission-dominated gas exchange of elemental mercury vapor over natural surfaces in China. *Atmos. Chem. Phys.* 16, 11125–11143 (2016).
- Paige Wright, L., Zhang, L. & Marsik, F. J. Overview of mercury dry deposition, litterfall, and throughfall studies. *Atmos. Chem. Phys.* 16, 13399–13416 (2016).

- Wang, X. et al. Underestimated sink of atmospheric mercury in a deglaciated forest chronosequence. *Environ. Sci. Technol.* 54, 8083–8093 (2020).
- Choi, H. D., Sharac, T. J. & Holsen, T. M. Mercury deposition in the Adirondacks: a comparison between precipitation and throughfall. *Atmos. Environ.* 42, 1818–1827 (2008).
- Pacyna, J. M. et al. Current and future levels of mercury atmospheric pollution on a global scale. *Atmos. Chem. Phys.* 16, 12495–12511 (2016).
- Wang, Y., Wu, P. & Zhang, Y. Climate-driven changes of global marine mercury cycles in 2100. Proc. Natl Acad. Sci. USA 120, e2202488120 (2023).
- Obrist, D. et al. A review of global environmental mercury processes in response to human and natural perturbations: changes of emissions, climate, and land use. *Ambio* 47, 116–140 (2018).
- 64. Christian, J. I. et al. Global distribution, trends, and drivers of flash drought occurrence. *Nat. Commun.* **12**, 6330 (2021).
- Shekhar, A., Hörtnagl, L., Buchmann, N. & Gharun, M. Long-term changes in forest response to extreme atmospheric dryness. *Glob. Change Biol.* **29**, 5379–5396 (2023).
- Allen, C. D., Breshears, D. D. & McDowell, N. G. On underestimation of global vulnerability to tree mortality and forest die-off from hotter drought in the Anthropocene. *Ecosphere* 6, art129 (2015).
- Zhao, M. & Running, S. W. Drought-induced reduction in global terrestrial net primary production from 2000 through 2009. *Sci*ence **329**, 940–943 (2010).
- Sun, T. et al. Mercury transport, transformation and mass balance on a perspective of hydrological processes in a subtropical forest of China. *Environ. Pollut.* 254, 113065 (2019).
- Zhang, Y., Soerensen, A. L., Schartup, A. T. & Sunderland, E. M. A global model for methylmercury formation and uptake at the base of marine food webs. *Glob. Biogeochem. Cycles* 34, e2019GB006348 (2020).
- Sunderland, E. M. & Mason, R. P. Human impacts on open ocean mercury concentrations. *Glob. Biogeochem. Cycles* https://doi. org/10.1029/2006GB002876 (2007).
- DiMento, B. P. & Mason, R. P. Factors controlling the photochemical degradation of methylmercury in coastal and oceanic waters. *Mar. Chem.* 196, 116–125 (2017).
- Rafaj, P., Bertok, I., Cofala, J. & Schöpp, W. Scenarios of global mercury emissions from anthropogenic sources. *Atmos. Environ.* 79, 472–479 (2013).
- Brocza, F. M., Rafaj, P., Sander, R., Wagner, F. & Jones, J. M. Global scenarios of anthropogenic mercury emissions. Preprint at EGUsphere https://doi.org/10.5194/egusphere-2024-41 (2024).
- Zhang, Y. et al. An updated global mercury budget from a coupled atmosphere-land-ocean model: 40% more re-emissions buffer the effect of primary emission reductions. One Earth 6, 316–325 (2023).
- Zhang, P. & Zhang, Y. Earth system modeling of mercury using CESM2 – Part 1: Atmospheric model CAM6-Chem/Hg v1.0. *Geosci. Model Dev.* 15, 3587–3601 (2022).
- Zhang, Y. et al. Observed decrease in atmospheric mercury explained by global decline in anthropogenic emissions. *Proc. Natl Acad. Sci. USA* **113**, 526–531 (2016).
- Khan, T. R., Obrist, D., Agnan, Y., Selin, N. E. & Perlinger, J. A. Atmosphere-terrestrial exchange of gaseous elemental mercury: parameterization improvement through direct comparison with measured ecosystem fluxes. *Environ. Sci. Process. Impacts* 21, 1699–1712 (2019).
- Wesely, M. L. Parameterization of surface resistances to gaseous dry deposition in regional-scale numerical models. *Atmos. Environ.* 23, 1293–1304 (1989).
- Lawrence, D. M., Koven, C. D., Swenson, S. C., Riley, W. J. & Slater, A. G. Permafrost thaw and resulting soil moisture changes regulate

projected high-latitude CO_2 and CH4 emissions. *Environ. Res. Lett.* **10**, 94011 (2015).

- Tang, B., Chen, J., Wang, Z., Qin, P. & Zhang, X. Mercury accumulation response of rice plant (*Oryza sativa* L.) to elevated atmospheric mercury and carbon dioxide. *Ecotoxicol. Environ. Saf.* 224, 11628 (2021).
- Stamenkovic, J. & Gustin, M. S. Nonstomatal versus stomatal uptake of atmospheric mercury. *Environ. Sci. Technol.* 43, 1367–1372 (2009).
- Millhollen, A. G., Obrist, D. & Gustin, M. S. Mercury accumulation in grass and forb species as a function of atmospheric carbon dioxide concentrations and mercury exposures in air and soil. *Chemosphere* 65, 889–897 (2006).
- 83. Natali, S. M. et al. Increased mercury in forest soils under elevated carbon dioxide. *Oecologia* **158**, 343–354 (2008).
- Benjamin, D. et al. Plant-soil distribution of potentially toxic elements in response to elevated atmospheric CO₂. *Environ. Sci. Technol.* 45, 2570–2574 (2011).
- Demers, J. D., Blum, J. D. & Zak, D. R. Mercury isotopes in a forested ecosystem: implications for air-surface exchange dynamics and the global mercury cycle. *Glob. Biogeochem. Cycles* 27, 222–238 (2013).
- 86. Gu, B. et al. Cost-effective mitigation of nitrogen pollution from global croplands. *Nature* **613**, 77–84 (2023).

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Competing interests

The authors declare no competing interests.

Additional information

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